2	Treatment with Copper-Containing-Waste
3	
4	Abdul Kadir Salam* dan Kiat Ginanjar
5	Faculty Staff* and Alumnus
6	Department of Soil Science University of Lampung
7	Bandar Lampung Indonesia 35145
8	
9	*Corresponding Author
10	axauam@gmail.com
11	Fax. 0721-8011325
12	

1 Abstract

Copper is reported to be retained in soils for a quite long time particularly in soil treated with some amendments. This research was intended to evaluate the soil labile fractions of $Cu \pm 10$ years after treatments with Cu-containing industrial waste, lime, and cassava-leaf compost. Soil samples were taken from topsoils and subsoils of ± 10 years old experimental plots set in 1998 and factorially treated with a metal-spoon industrial waste at 0, 15, and 60 ton ha⁻¹, lime at 0 and 5 ton ha⁻¹, and cassava-leaf compost at 0 and 5 ton ha⁻¹. The measured Cu labile fractions were compared to those in soils sampled ± 1.5 years and ± 3 years after treatments. The results showed that the soil Cu labile fractions in waste treated soil were higher than those in the control treatments even though their concentration decreased with the years of sampling. Lime showed a decreasing effect on soil labile Cu fractions, but the effect decreased with the years of sampling. Effect of cassava-leaf compost was in general not observed ± 10 years after treatment.

Keywords: Cassava-Leaf Compost, Copper, Heavy Metals, Industrial Waste, Lime

16 Introduction

Heavy metals are by definition the elements with relatively high atomic density of > 6 g cm⁻³ or specific gravity of > 5 g cm². This category of elements is one of the research major focuses in soil science because heavy metals are very important in the environment. Some heavy metals are needed by plants and animals at relatively small amounts. For example, Cu and Zn are considered microelements for plant growth; therefore, they are needed by plants, even though only at relatively small amounts. On the contrary, some of the elements such as Cd and Pb are not needed by the living things. Both categories of these elements show similarity, they are toxic to the living things at concentrations relatively higher than their critical levels. For examples, the total concentrations of

1 Cu in soils must not exceed 60 mg kg⁻¹; that of Zn must be below 70 mg kg⁻¹; while Cd and Pb must 2 be < 9 and 100 mg kg⁻¹, respectively (Ross 1994a). These values are toxic boundaries for plants.

However, the concentrations of heavy metals in soils are subjected to natural and anthropogenic sources (Juracek and Ziegler 2006; Biasioli et al. 2007; Benke et al. 2008; Berenguer et al. 2008; Lin et al. 2008; Cakmak et al. 2010; Alloway 2012; Tu et al. 2012; Kargar et al. 2013; Adams et al. 2014). Berenguer et al. (2008) reported that long-term use of liquid swine manure significantly elevated the concentrations of Cu and Zn in soils. Long-term application of cattle manure significantly increased the soil total and EDTA-extractable Cu and Zn (Benke et al. 2008). The concentrations of Cu, Ni, Hg, Pb, Cd, and As in sediment samples at electronic waste recycling plant at Guiyu, Guangdong, China significantly increased above the reference background concentrations (Guo et al. 2009). Therefore, heavy metals in the soil environment must be one of our concern.

Heavy metals in soil environment are distributed in several foms with decreasing bioavailability: free ions, complex ions, exchangeable forms, precipitates, and minerals (Salam 2017). Among these, free ions and to some extent complex ions and exchangeable forms, are the most mobile and potential to affect the living things because these forms are directly related to plant root absorption and heavy metal toxicities (Salam 2001; Daoust et al. 2006; Salam 2017). Therefore, the behaviors of these labile fractions of heavy metals in soils must be understood. This may include their behaviours related to the dynamics of some soil key properties such as soil cation exchange capacity (CEC), pH, and reaction time.

Soil adsorptive capacity – part of it is expressed as CEC – is by no means the most important soil property for heavy metal immobilization. This property may be improved by addition of ameliorants. For examples, some researchers empoyed organic matters/biosolids (Salam et al. 2001 2005 2017; Tokunaga et al. 2003; Brown et al. 2004; Stehouwer et al. 2006; Schroder et al. 2008; Brown et al. 2009; Kukier et al. 2010; Mamindy-Pajany et al. 2014; Tang et al. 2014; Pukalchik et al. 2017) or P fertilizers such as TSP and rock phosphates (Brown et al. 2004; Kilgour et al. 2008;

1 Moseley et al. 2008). Organic matters are believed to give functional groups such as phenolic and

carboxyls in significant amounts and to increase soil CEC (Ross 1994b; Parfitt et al. 1995; Rodella

et al. 1995; Salam 2001). Lin et al. (2008) reported that Cu tended to be preferentially retained by

4 Fe-oxides and organic matters. Some other ameliorants were also developed to immobilize heavy

metals by precipitation, for examples by employing Na₂HPO₄, hidroxyapatite, or rock phosphates

(Ma et al. 1990 1994 1995; Rabinowitz 1993; Ruby et al. 1994).

As repeatedly reported by several researchers (Salam 1999 2001; Adams et al. 2004; Bang and Hesterberg 2004; Quaghebeur et al. 2005; He et al. 2006; Brown et al. 2009; Bolan et al. 2014; Malinowska 2017), the soil adsorptive capacity or CEC is positively related to the changes in soil pH. In general, soil CEC increases with the increase in pH of soils with variable charges. The increase is due to dehydogenation of soil particle surfaces in the presence of increased concentration of OH. The increase in soil CEC increases the adsorption of heavy metals. For example, it was reported that amendment of soil contaminated with industrial waste with 5 ton ha-1 CaCO₃ significantly increased the soil pH dan decreased the labile heavy metals extracted by chemical extractors (Salam 2000 2001 2017). This behavior then increases the residence time of heavy metals in soils and decreases their availabilities to plants and mobilities in soil environment. Daoust et al. (2006) lately reported that Cu partitioning and its toxicity were significantly affected by pH greater than by organic matter and clay content.

This research was intended to evaluate the soil labile fraction of $Cu \pm 10$ years after soil treatments with Cu-containing industrial waste, lime, and cassava-leaf compost.

Materials and Methods

Soil samples were taken from topsoils and subsoils in the experimental plots set in July 1998, located in Sidosari, Natar, South Lampung. Treatments were previously set in a randomized block design, consisted of 3 factors: industrial waste of metal spoon, lime as dolomite, and cassava-leaf compost, with 3 replications. Industrial waste was metal-spoon industry of PT Star Metal Wares

1 Jakarta, containing high Cu and Zn. Some chemical properties of the industrial waste were pH 7.30,

2 Pb 2.44 mg kg⁻¹, Cd 0.12 mg kg⁻¹, Cu 754 mg kg⁻¹, and Zn 44.6 mg kg⁻¹ (Salam et al. 2005).

Industrial waste was given and thoroughly plowed to 30 cm soil depth at rates: 0, 15, and 60 ton ha⁻¹. Lime and cassava-leaf compost were both given and thoroughly mixed to the same soil depth one week later at 0 and 5 ton ha⁻¹. Lime was (CaMg(CO₃)₂). Casava-leaf compost was prepared as reported previously by Salam (2001). Rates of industrial waste, lime, and cassava-leaf compost were selected through a series if preliminary experiments reported previously (Salam 2000). Each plot measured 4.5 m long and 4 m wide, 50 cm apart between plots and 100 cm apart between blocks. Complete experimental treatments are listed in Table 1. The experimental plots were since 1998 planted with corn, dryland paddy, and left bare in between.

Soil sampling was conducted in 15 February 2009 (± 10 years after treatment). Composite topsoil (0-15 cm) and subsoil (15-30 cm) samples were taken diagonally from 5 points in each plot. Soil samples were air-dried, ground to pass a 2-mm-sieve, and mixed thoroughly before analysis. Analysis included soil Cu labile fraction with DTPA method (Baker and Amacher 1987) and soil pH with a pH electrode. This data was compared to those of ± 1.5 years after treatment (Amirulloh 2000) and ± 3 years after treatment (Prihatin 2002).

Insert Table 1

Results and Discussion

Copper was chemically retained in soil body for more than 10 years since soil treatment with Cucontaining industrial waste in July 1998. Analysis of variance (Anova) shows that the industrial waste significantly enhanced the soil labile factions of Cu (Table 2). Test of Least Significant Difference (LSD-Test) also shows the significance of this effect, particularly by the addition of the 1 industrial waste of 60 ton ha⁻¹ (Table 3). The concentration of Cu at the addition of 60 ton ha⁻¹ was

3.93 times compared to that at Control at 0 - 15 cm, while that at 15- 30 cm was 2.55 times. The

3 addition of industrial waste at 15 ton ha⁻¹ also shows tendencies to increase the concentrations of Cu;

1.82 times and 1.49 times at 0 - 15 and 15 - 30 cm, respectively. The increases were clearly

attributed to the fact that the industrial waste contained relatively high amount of Cu (Salam et al.

6 2005).

Insert Table 2

Insert Table 3

However, the concentrations of soil labile Cu in general decreased significantly with year of sampling, particularly at the addition level of 60 ton ha⁻¹. The concentrations of labile Cu at ± 1.5 years were higher than those at ± 3 years and those at ± 3 years were higher than those at ± 10 years (Figs. 1, 2, and 3). The relative concentrations of labile Cu in topsoils treated with 60 ton waste ha⁻¹ and 5 ton lime ha⁻¹ decreased from 0.67 at ± 1.5 years to 0.54 at ± 3 years and to 0.39 at ± 10 years. The relative concentration of Cu is expressed as Cu/Cu₀-1.5 i.e. the ratio of the concentration of Cu to that of the Control (No Compost and No Lime) at ± 1.5 years. The relative concentrations of Cu at waste level of 60 ton ha⁻¹ are listed in Tabel 4.

Insert Table 4

The decreasing trend in the concentration of Cu for the last ± 10 years was due to several possibilities: (1) enhanced retainment by soil adsorptive surfaces or by precipitation, (2) absorbed by plant roots, and (3) leached by percolating water to subsoil or ground water, as described in Fig. 4.

Based on the observation, the first possibility was unlikely. First, at ± 10 years after treatment, the effect of lime had decreased significantly and, based on Anova, the effect on decreasing Cu

labile fractions was not significant (Table 1). Second, the interaction of Industrial Waste of 15 ton 1 ha⁻¹ with lime tended to decrease the concentration of soil labile Cu. However, the effect was not 2 significant; and, on the contrary, the interaction with Industrial Waste of 60 ton ha⁻¹ in fact increased 3 4 the concentration of soil labile Cu. The interaction of Industrial Waste and Cassava-Leaf Compost 5 also significantly decreased the soil labile Cu. However, this material was easily decomposed. 6 Therefore, the logical possibility causing the decrease in the concentrations of labile Cu with time 7 was Cu absorption by plant roots or Cu translocation into subsoil or ground water (Fig. 4). The 8 complexation of Cu with soil humic substances may have mobilized Cu ions so that it was more 9 easily transported and/or absorbed by plant roots. In fact, Cu is more easily complexed by organic

11

10

12 Insert Fig. 1

substances.

- 13 Insert Fig. 2
- 14 Insert Fig. 3
- 15 Insert Fig. 4

16

17

18

19

20

Summary and Conclusion

Copper of industrial-waste-origin was retained in soil for more than 10 years. The concentration of labile Cu decreased with year of soil sampling. However, part of the Cu from industrial waste was translocated, probably absorbed by plant roots and/or moved into subsoil and/or ground water through soil body.

21

22

Acknowledgments

- 23 This research was supported by the research funding provided by The Directorate General of Higher
- 24 Education The Department of National Education through "Hibah Kompetitif Penelitian Sesuai
- 25 **Prioritas Nasional**" in the Fiscal Year of 2009.

3

Literature	Citea

- 5 Adams ML, FJ Zhao, SP McGrath, FA Nicholson and BJ Chambers. 2004. Predicting cadmium
- 6 concentrations in wheat and barley grain using soil properties. J Environ Qual 33: 532-541.
- 7 Alloway BJ. 2012. Sources of heavy metals and metalloids in soils. Heavy Metals Soils 22: 11-50.
- 8 Amirulloh. 2000. Tinggalan Unsur Mikro Kelompok Logam Berat Asal Limbah Industri dalam
- 9 Tanah yang Telah Diperlakukan dengan Kapur dan Kompos Daun Singkong. [Script].
- 10 University of Lampung, Bandar Lampung. (in Indonesian).
- Baker DE and MC Amacher. 1987. Nickel, copper, zinc, and cadmium. In: AL Page, RH Miller,
- and DR Keeney (eds.). Methods of Soil Analysis Part 2 Chemical and Microbiological
- Properties. 2nd ed. SSSA Inc., Madison, pp. 323-336.
- 14 Bang JS and D Hesterberg. 2004. Dissolution of trace element contaminants from two coastal plain
- soils as affected by pH. J Environ Qual 33: 891-901.
- Benke MB, SP Indraratne, X Hao, C Chang and TB Goh. 2008. Trace element changes in soil after
- long-term cattle manure applications. J Environ Qual 37: 798-807.
- 18 Berenguer SC, F Santeveri, J Boixadera and J Lloveras. 2008. Copper and zinc soil accumulation
- and plant concentration in irrigated maize fertilized with liquid swine manure. Agron J 100:
- 20 1056-1061.
- 21 Biasioli M, H Greman, T Kralj, F Madrid, E Diaz-Barrientos and F Ajmone-Marsan. 2007.
- 22 Potentially toxic elements contamination in urban soils. J Environ Qual 36: 70-79.

- Bolan N, A Kunhikhrisnan, R Thangarajan, J Kumpiene, J Park, T Makino, MB Kirkham and K
- 2 Scheckel. 2013. Remediation of heavy metal(loid)s contaminated soils to mobilize or to
- 3 immobilize? J Hazardous Materials 266: 141-166.
- 4 Brown S, RF Chaney, J Hallfrisch, JA Ryan and WR Berti. 2004. *In situ* soil treatments to reduce
- 5 the phyto- and bioavailability of lead, zinc, and cadmium. J Environ Qual 33: 522-531.
- 6 Brown S, A Svendsen and C Henry. 2009. Restoration of high zinc and lead tailings with municipal
- 7 biosolids and lime: a field study. J Environ Qual 38: 2189-2197.
- 8 Cakmak D, E Saljnikov, V Mrvic, M Jakovljevic, Z Marjanovic, B Sikiric and S Maksimovic. 2010.
- 9 Soil properties and trace elements contents following 40 years of phosphate fertilization. J
- 10 Environ Qual 39: 541-547.
- Daoust CM, C Bastien and L Deschenes. 2006. Influence of soil properties and aging on the
- toxicity of copper on compost worm and barley. J Environ Qual 35: 558-567.
- 13 Ginanjar K. 2009. Fraksi Labil Tembaga dan Seng dalam Tanah pada 10 Tahun Setelah Perlakuan
- dengan Limbah Industri. [Script]. University of Lampung, Bandar Lampung. (in Indonesian).
- 15 GuoY, C Huang, H Zhang and Q Dong. 2009. Heavy metal contamination from electronic waste
- recycling at Guiyu, Southern China. J Environ Qual 38: 1617-1626.
- He ZJ, M Zhang, XE Yang and PJ Stoffella. 2006. Release behavior of copper and zinc from sandy
- 18 soils. Soil Sci Soc Am J 70: 1699-1707.
- 19 Juracek KE and AC Ziegler. 2006. The legacy of leaded gasoline in bottom sediment of small rural
- 20 reservoirs. J Environ Qual 35: 2092-2102.
- 21 Kargar M, P Jutras, OG Clark, WH Hendershot and SO Prasher. 2013. Trace metal contamination
- influenced by land use, soil age, and organic matter in Montreal tree pit soil. J Environ Qual
- 23 42(5): 1527-1533.

- 1 Kilgour DW, RB Moseley, MO Barnett, KS Savage and PM Jardine. 2008. Potential negative
- 2 consequences of adding phosphorus-based fertilizers to immobilize lead in soils. J Environ
- 3 Qual 37: 1733-1740.
- 4 Kukier U, RL Chaney, JA Ryan, WL Daniels, RH Dowdy and TC Granato. 2010. Phytoavailability
- of cadmium in long-term biosolids-amended soils. J Environ Qual 39: 519-530.
- 6 Lin C, I Negev, G Eshel and A. Banin. 2008. In situ accumulation of copper, chromium, nickel, and
- 7 zinc in soils used for long-term waste water reclamation. J Environ Qual 37: 1477-1487.
- 8 Malinowska E. 2017. The effect of liming and sewage sludge application on heavy metal speciation
- 9 in soil. Bull Environ Contam Toxicol 98(1): 105-112.
- Mamindy-Pajany Y, S Sayen, JFW Mosselmans and E Guillon. 2014. Copper, nickel, and zinc
- speciation in biosolid-amended soil: pH adsorption edge, μ -XRF and μ -XANES investigations.
- 12 Environ Sci Technol 48(13): 7227-7244.
- 13 Ma QY, SJ Traina and TJ Logan. 1990. In situ lead immobilization by apatite. Environ Sci
- 14 Technol 27: 1803-1810.
- 15 Ma QY, SJ Traina, TJ Logan and JA Ryan. 1994. Effects of aqueous Al, Cd, Cu, Fe(II), Ni, and Zn
- on Pb immobilization by hydroxyapatite. Environ Sci Technol 28: 1219-1228.
- 17 Ma QY, TJ Logan and SJ Traina. 1995. Lead immobilization from aqueous solution and
- contaminated soils using phosphate rocks. Environ Sci Technol 29: 1118-1126.
- 19 Moseley RA, MO Barnett, MA Stewart, TL Mehlhorn, PA Jardine, M Ginder-Vogel and S Fendorf.
- 20 2008. Decreasing lead bioaccessibility in industrial and firing range soils with phosphate-
- based amendments. J Environ Qual 37:2 116-2124.
- 22 Parfitt RL, DJ Giltrap and J.S. Whitton. 1995. Contribution of organic matter and clay minerals to
- the cation exchange capacity of soils. Commun. Soil Sci Plant Anal 26: 1343-1355.

- 1 Prihatin F. 2002. Pengaruh Tinggalan Limbah Industri, Kapur, dan Kompos Daun Singkong
- 2 terhadap Ketersediaan Cu dan Zn pada Beberapa Kedalaman Tanah Ultisol Sidosari Natar
- 3 Lampung Selatan. [Script]. University of Lampung, Bandar Lampung. (in Indonesian).
- 4 Pukalchik M, M Panova, M Karpukhin, O Yakimenko, K Kydralieva and V Terekhova. 2017.
- 5 Using humic products as amendment to restore Zn and Pb polluted soil: a case study using
- 6 rapid screening phytotest endpoint. J Soils Sediments. https://doi.org/10.1007/s11368-017-
- 7 1841-y.
- 8 Quaghebeur M, A Rate, Z Rengel and C Hinz. 2005. Desorption kinetics of arsenate from kaolinite
- 9 as influenced by pH. J Environ Qual 34: 479-486.
- 10 Rabinowitz MB. 1993. Modifying soil lead bioavailability by phosphate addition. Bull Environ
- 11 Contam Toxicol 51: 438-444.
- Ranur S. 2009. Status Pergerakan Tembaga dan Seng dalam Tubuh Tanah di Desa Sidosari Natar,
- Lampung Selatan, pada 10 Tahun Setelah Perlakuan dengan Limbah Industri. [Script].
- 14 University of Lampung, Bandar Lampung. (in Indonesian).
- 15 Rodella AA, KR Fisher and JC Alcarde. 1995. Cation exchange capacity of an acid soil as
- influenced by different sources of organic matter. Commun Soil Sci Plant Anal 26: 2691-
- 17 2967.
- Ross SM 1994a. Sources and forms of potentially toxic metals in soil-plant systems. In: SM Ross
- 19 (ed.). Toxic Metals in Soil-Plant Systems. John Willey & Sons, Ltd., New York. pp. 3-25.
- 20 Ross SM 1994b. Retention, transformation, and mobility of toxic metals in soils. In S.M. Ross
- 21 (ed.). Toxic Metals in Soil-Plant Systems. John Willey & Sons, Ltd., New York. pp.63-152.
- 22 Ruby MV, A Davis and A Nicholson. 1994. In situ formation of lead phosphates in soils as a
- 23 method to immobilize lead. Environ Sci Technol 28: 646-654.

- 1 Salam AK. 2000. A four year study on the effects of manipulated soil pH and organic matter
- 2 contents on availabilities of industrial-waste-origin heavy metals in tropical soils. J Tropical
- 3 Soils 11: 31-46.
- 4 Salam AK. 2001. Manajemen Fraksi Labil Logam Berat di dalam Tanah untuk Pertanian dan
- 5 Lingkungan. University of Lampung Press, Bandar Lampung. 62p. (in Indonesian).
- 6 Salam AK, S Bakrie and F Prihatin. 2005. Depth-wise distribution of extracted Cu and Zn in
- 7 cultivated field-plots three years after treatment with a Cu- and Zn-containing waste, lime, and
- 8 cassava-leaf compost. J Tropical Soils 11(1): 9-14.
- 9 Salam AK. 2017. Management of Heavy Metals in Tropical Soil Environment. Global Madani
- 10 Press, Bandar Lampung. 257p.
- 11 Schroder JL, H Zhang, D Zhou, N Basta, WR Raun, ME Payton and A Zazulak. 2008. The effect of
- long-term annual application of biosolids on soil properties, phosphorrus, and metals. Soil Sci
- 13 Soc Am.J 72: 73-82.
- 14 Stehouwer R, RL Day and KE Macneal. 2006. Nutrient and trace element leaching following mine
- reclamation with biosolids. J Environ Qual 35: 1118-1126.
- 16 Tang WW, GM Zeng, JL Gong, J Liang, P Xu, C Zhang and B.B. Huang. 2014. Impact of
- humic/fulvic acid on the removal of heavy metals from aqueous solution using nanomaterials:
- 18 a review. Sci Total Environ 468-469: 1014-1027.
- 19 Tokunaga TK, J Wan, MK Firestone, TC Hazen, KR Olson, DJ Herman, SR Sutton and A.
- 20 Lanzirotti. 2003. In situ reduction of chromium (VI) in heavily contaminated soil through
- organic carbon amendment. J Environ Qual 32: 1641-1649.
- 22 Tu CL, TB He, CQ Liu, and XH. Lu. 2012. Effects of land use and parent materials on trace
- elements accumulation in topsoil. J Environ Qual., 42(1):103-110.

1 Table 1. The existing treatment units in the experimental plots at Sidosari, Natar, Lampung.

Waste/W	Lime /L –	Comp	post/C
vv aste/ vv	Line/L —	C_0	C_1
	L_0	$W_0L_0C_0$	$W_0L_0C_1$
\mathbf{W}_0	L1	$W_0L_1C_0$	$W_0L_1C_1$
	L_0	$W_1L_0C_0$	$W_1L_0C_1$
\mathbf{W}_1	L_1	$W_1L_1C_0$	$W_1L_1C_1$
	L_0	$W_2L_0C_0$	$W_2L_0C_1$
\mathbf{W}_2	L_1	$W_2L_1C_0$	$W_2L_1C_1$

Notes:

2

3 4 5 6 $W = Waste (W_0 0, W_1 15, and W_2 60 ton ha^{-1}), L = Lime (L_0 0 and L_1 5 ton ha^{-1}); and C = Compost (C_0 0 and C_1 5 ton ha^{-1}).$

Table 2. Analysis of variance of the changes in labile Cu concentration in a tropical soil treated with Cu-containing industrial waste, lime, and cassava-leaf compost after a period of time since treatment (Transf \sqrt{x}) (Ginanjar 2009).

	Time (years)					
Treatment	± 1.5 ¹⁾		$\pm 3^{2)}$		± 10	
	0-15 cm	15-30 cm	0-15 cm	15-30 cm	0-15 cm	15-30 cm
W	**	**	**	**	**	**
С	ns	ns	ns	**	ns	ns
L	**	ns	ns	ns	ns	ns
WxC	ns	ns	ns	**	**	**
WxL	ns	ns	**	ns	**	**
CxL	ns	ns	ns	**	ns	ns
WxCxL	ns	ns	**	**	ns	**

Notes: W = Waste; C = Cassava-Leaf Compost; L = Lime; * = Significant at 5%; ** = Significant at 1%; ns = Not Significant at 5% and 1%; ¹⁾ Amirulloh (2000); ²⁾ Prihatin (2002).

Table 3. The effect of waste on the concentrations of Cu labile fraction in a tropical soil ± 10 years after treatment with industrial waste (Trans \sqrt{x}) (Ginanjar 2009).

Waste Levels	Soil Depth (cm)		
(t ha ⁻¹)	0-15	15-30	
	\(\sqrt{mg / kg} \)		
0	2,68 a	1,78 a	
15	3,62 a	2,17 a	
60	5,31 b	2,84 b	
LSD 5%	1,01	0,64	

Notes: Figures in one column with the same characters are not significantly different at 5%.

1 Table 4. The relative concentrations of Cu (Cu/Cu $_0$ -1.5) at waste level of 60 ton ha $^{\text{-}1}$.

No.	Treatment Unit	± 1.5 years	± 3 years	± 10 years
1	No C, No L	1.00	0.77	0.19
2	No C, With L	0.67	0.54	0.39
3	With C, No L	0.80	0.60	0.21
4	With C, With L	0.55	0.64	0.17

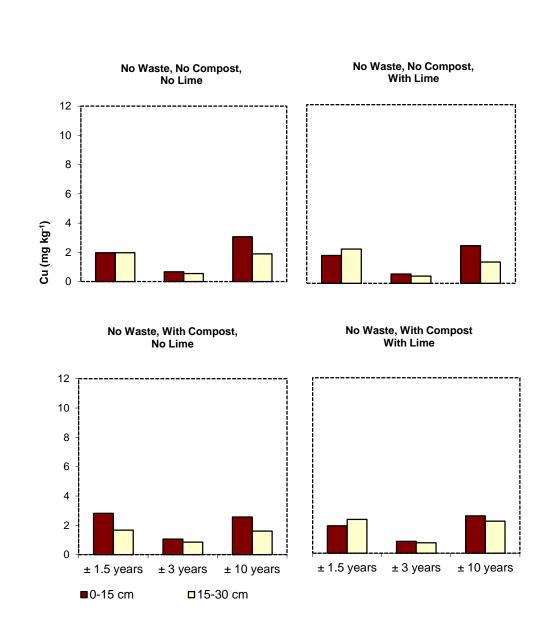


Fig 1. The effect of industrial waste, lime, and cassava-leaf compost on the concentration of a tropical soil labile fraction of Cu (Trans \sqrt{x}) at waste level 0 ton ha⁻¹ (Compost 5 ton ha⁻¹ and Lime 5 ton ha⁻¹). Amirullah (2000); Prihatin (2002) (Ginanjar 2009).

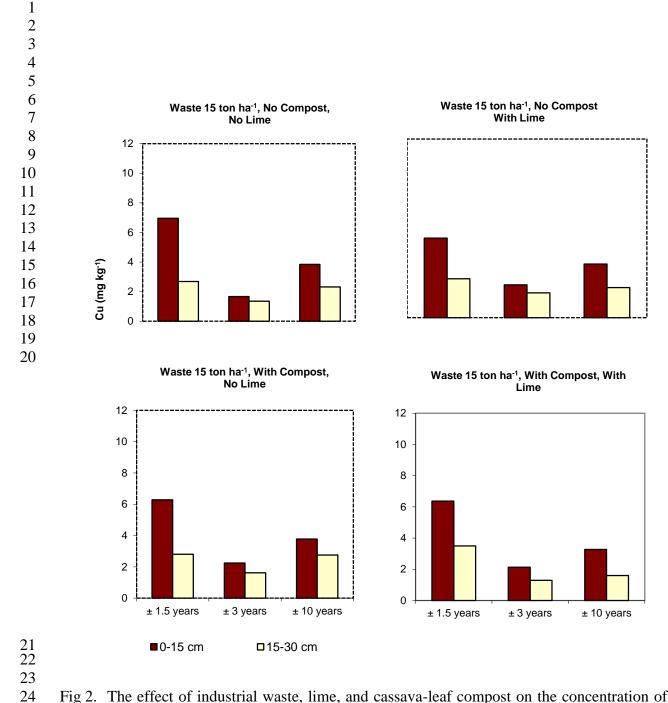


Fig 2. The effect of industrial waste, lime, and cassava-leaf compost on the concentration of labile fraction of Cu (Transf√x) at waste level of 15 ton ha⁻¹ (Compost 5 ton ha⁻¹ and Lime 5 ton ha⁻¹). ¹¹) Amirullah (2000); ²¹)Prihatin (2002) (Ginanjar 2009).

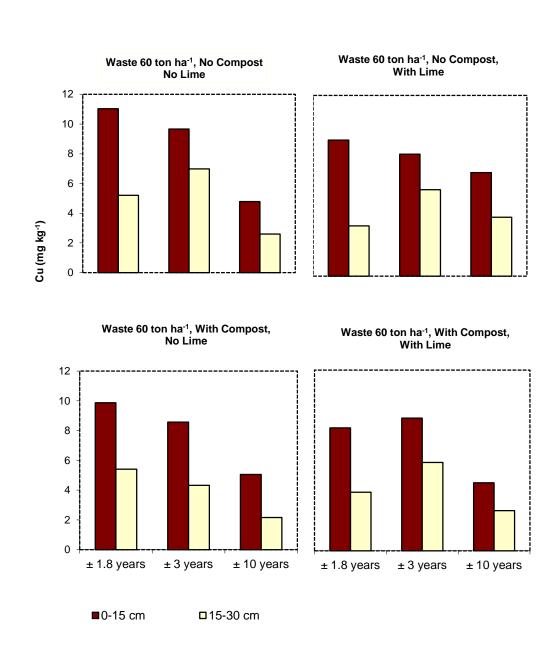


Fig 3. The effect of industrial waste, lime, and cassava-leaf compost on the concentration of a tropical soil labile fraction of Cu (Transf \sqrt{x}) at waste level of 60 ton ha⁻¹ (Compost 5 ton ha⁻¹ and Lime 5 ton ha⁻¹). Amirullah (2000); Prihatin (2002) (Ginanjar 2009).

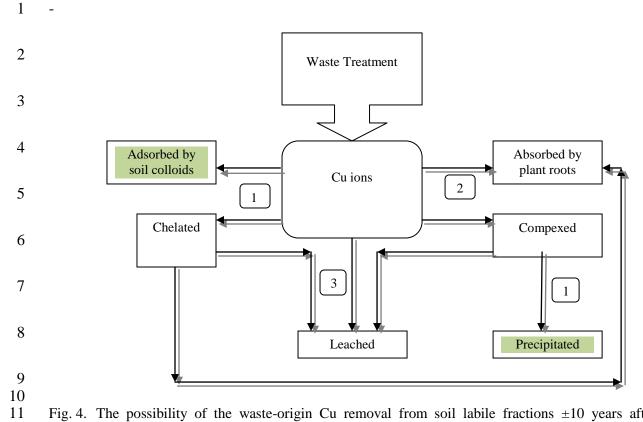


Fig. 4. The possibility of the waste-origin Cu removal from soil labile fractions ± 10 years after waste treatment (1 – Adsorption or precipitation, 2 – Absorption by plant roots, 3 – leaching).